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# Effects of acute and chronic temperature changes on the functional responses of the dogfish *Scyliorhinus canicula* (Linnaeus, 1758) towards amphipod prey *Echinogammarus marinus* (Leach, 1815)

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**Abstract** Predation is a strong driver of population dynamics and community structure and it is essential to reliably quantify and predict predation impacts on prey populations in a changing thermal landscape. Here, we used comparative functional response analyses to assess how predator-prey interactions between dogfish and invertebrate prey change under different warming scenarios. The Functional Response Type, attack rate, handling time and maximum feeding rate estimates were calculated for *Scyliorhinus canicula* preying upon *Echinogammarus marinus* under temperatures of 11.3 °C and 16.3 °C, which represent both the potential daily variation and predicted higher summer temperatures within Strangford Lough, N. Ireland. A two x two design of “Predator Acclimated”, “Prey Acclimated”, “Both Acclimated”, and “Both Unacclimated” was implemented to test functional responses to temperature rise. Attack rate was higher at 11.3 °C than at 16.3 °C, but handling time was lower and maximum feeding rates were higher at 16.3 °C. Non-acclimated predators had similar maximum feeding rate towards non-acclimated and acclimated prey, whereas acclimated predators had significantly higher maximum feeding rates towards acclimated prey as compared to non-acclimated prey. Results suggests that the predator attack rate is decreased by increasing temperature but

when both predator and prey are acclimated the shorter handling times considerably increase predator impact. The functional response of the fish changed from Type II to Type III with an increase in temperature, except when only the prey were acclimated. This change from population destabilizing Type II to more stabilizing Type III could confer protection to prey at low densities but increase the maximum feeding rate by *Scyliorhinus canicula* in the future. However, predator movement between different thermal regimes may maintain a Type II response, albeit with a lower maximum feeding rate. This has implications for the way the increasing population *Scyliorhinus canicula* in the Irish Sea may exploit valuable fisheries stocks in the future.

**Keywords** Thermal responses · Functional response · Feeding ecology · Climate change

## Introduction

Temperature is a driver of interspecific interactions through mediating metabolism (Arrhenius 1889; Brown et al. 2004) and therefore locomotion (Dell et al. 2011; Dell et al. 2014), attack speed (Nowicki et al. 2012), feeding rates (Iacarella et al. 2015), and growth (Savage et al. 2004; Pörtner and Knust 2007). Therefore, temperature has a direct effect upon energy flux through an ecosystem and the population dynamics within it (Woodward et al. 2010; Stevnbak et al. 2012; Cahill et al. 2013). Thermal responses are understood to be an underlying mechanism behind predator – prey

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dynamics (Berlow et al. 2009; Rall et al. 2010; Englund et al. 2011), but predicting the outcome of such interactions under future abiotic scenarios is highly problematic (Le Quesne and Pinnegar 2011). This is partly due to multiple layers of real world complexities such as habitat heterogeneity (MacNeil et al. 2004; Ferner et al. 2009; Alexander et al. 2015; Barrios-O'Neill et al. 2015), light (Koski and Johnson 2002), multiple predator effects (Lang et al. 2012; Alexander et al. 2013; Wasserman et al. 2016a). The apparent species specificity of thermal dynamics poses an additional layer of difficulty in response prediction due to the metabolism of different species having varying sensitivity to temperature effects and therefore producing differential responses (Lang et al. 2012). It is thus difficult to derive accurate conclusions from thermal responses modeled across numerous systems (Grigaltchik et al. 2012), but it is widely agreed that temperature change alters ecological stability (Dell et al. 2014; Gilbert et al. 2014).

Predicting future responses to warming temperatures is complex due to systems being multifaceted in their response to temperature change. The effects of temperature change can be discernable on an individual level (Huey et al. 1999; Calosi et al. 2013; Vasseur et al. 2014), and on population and community levels (Le Quesne and Pinnegar 2011; Rall et al. 2012). Furthermore, thermal change may manifest as long-term chronic temperature increase or as acute temperature variations over short periods of time (Vasseur et al. 2014). While chronic warming allows time for adaption to a higher temperature, short-term temperature variation exposes organisms rapidly to potentially adverse conditions. Both scenarios are caused by ongoing climate change (Easterling et al. 1997, 2000) and can affect the life history and fitness of organisms (Yocom and Edsall 1974; Biro et al. 2010; Terblanche et al. 2010; Clusella-Trullas et al. 2011).

Prediction of consumer impact has been assessed through the use of comparative functional response analyses (Smout et al. 2010; Alexander et al. 2012; Alexander et al. 2013; Dick et al. 2013; Dick et al. 2014; Laverty et al. 2014; Wasserman et al. 2016a, b; Dick et al. 2017). This method has been pioneered for the prediction of invasive species impacts in comparison to trophically analogous native species (Dick et al. 2013; Paterson et al. 2015; Dick et al. 2017). It is particularly useful in assessing ecological impact under a number of abiotic contexts (e.g. Alexander et al. 2015). Functional response refers to the rate of consumption of a resource

with regards to resource density (Solomon 1949; Holling 1959). It is clear that species with a higher functional response (e.g., higher attack rates, lower handling times, higher asymptote) have a higher impact potential (Bollache et al. 2008; Haddaway et al. 2012; Dick et al. 2013). Functional responses are categorised into: the Type I rectilinear response, usually only found in filter feeders where resource consumption is not limited by handling time (Jeschke et al. 2004); the Type II inversely density dependent response with high prey consumption at low densities; and Type III sigmoidal, positively density dependent response whereupon prey have a low-density refuge (Oaten and Murdoch 1975). Within these response types, it is also possible to quantify the associated parameters of attack rate, handling time and maximum feeding rates (Juliano 2001; Dick et al. 2013). Functional response analysis therefore presents itself as a robust assessment method whereupon ecological impact can be explained and predicted across numerous biotic and abiotic contexts (Dick et al. 2014; Dick et al. 2017).

Warming oceans have the potential to alter interaction strengths between valuable commercial and non-commercial fisheries (Le Quesne and Pinnegar 2011). This has implications for maintaining biodiversity and economic value. As such there should be an emphasis on predicting and understanding the potential changes in species ecology under such climatic scenarios. This will pave the way for integrating results into management techniques well in advance so as to attempt to mitigate adverse change in community ecology and interaction strengths. We investigated predation by juvenile *Scyliorhinus canicula* (lesser-spotted dogfish) on amphipod prey *Echinogammarus marinus*. Both species are ubiquitous in temperate waters (Ellis and Shackley 1997; Duffy and Harvilicz 2001), resilient (Múrias et al. 1996; Revill et al. 2005; Leite et al. 2014) and pliable laboratory species (Sims 2003). These are ecologically relevant species as *E. marinus* has large impacts upon community structuring (Duffy and Hay 2000; Dick et al. 2005), and is an important prey species (Leite et al. 2014) that can additionally be considered a proxy for other small crustacean species and small fish due to strong swimming behaviours. *S. canicula* is a generalist benthic predator (Kaiser and Spencer 1994; Domi et al. 2005) with high abundances (Genner et al. 2010; Sguotti et al. 2016) that have been found to increase with increasing temperature (Sguotti et al. 2016). *S. canicula* presents as an ideal indicator species that is not affected by fishing

mortality, due to lack of fishing pressure and high discard survival rate (Revill et al. 2005).

The present study presents the functional responses of juvenile *Scyliorhinus canicula* preying on live *Echinogammarus marinus* under the ambient temperature of  $11 \pm 0.09$  °C and attempts to elucidate whether there is a change in the functional response when temperature is increased to  $16.3 \pm 0.20$  °C, a value chosen as it is in line with future Sea Surface Temperature (SST) increase predictions (Sokolov et al. 2009). Furthermore, we attempted to use comparative functional responses to understand whether predator or prey was more profoundly affected by the temperature increase. A  $2 \times 2$  factorial design was employed wherein both predator and prey were acclimated, just the predator was acclimated, just the prey were acclimated, and both predator and prey were unacclimated to 16.3 °C. This takes into account the effect of increased temperature on resource consumption and the movement of predators between different thermal regimes (Sims et al. 2001) in an attempt to predict possible changes in predator-prey interaction strength in the future.

## Methods

Thirty captive bred juvenile *Scyliorhinus canicula* ( $34.82 \pm 2.77$  SD cm total length) were kept in three holding tanks ( $120 \times 60 \times 60$  cm) at Queen's Marine Laboratory (QML Portaferry) and maintained on chopped mussel and mackerel ad lib. Three times a week. Trials were carried out from May–June 2016. The holding tanks were supplied with 11 °C UV filtered seawater pumped directly from Strangford Lough, Northern Ireland. *Echinogammarus marinus* were collected by searching under cobbles on 'Walter's Shore', Portaferry. These were kept in aerated 28 L boxes. *E. marinus* was fed with *Laminaria digitata* which was replaced once a week.

Experimental setup involved three leader tanks ( $120 \times 60 \times 60$  cm) heated with aquarium heaters. Three flow-through systems provided seawater into the leader tank to be heated, the water then flowed from the leader tank into six 60 L arenas. Airlines were fed directly into each leader tank to avoid confounds of low dissolved oxygen concentration (kept above 70% saturation DO). Acclimation to the desired temperature took place over three days as temperature was increased from ambient to 16.3 °C at a rate of 2 °C every two days, and then a

further two days at 16.3 °C. Ten lesser-spotted dogfish were kept in each leader tank during acclimation. *Echinogammarus marinus* were acclimated in 28-l boxes via the flow through system. The dogfish were fed ad lib. Throughout this time to avoid hyperphagy (Wootton and Ali 2001; Ali et al. 2003). Once the desired temperature was achieved, the water was held at  $16.3 \pm 0.20$  °C. The mean summer temperature in Strangford Lough is 15.5 °C (AFBI 2016). The raised temperature value reflects the potential daily variation in the summer and the 3–5 °C global (SST) increase predicted in the coming century (Sokolov et al. 2009).

A  $2 \times 2$  experiment design was implemented wherein Factor 1 was "Predator Treatment" with respect to non-acclimation/acclimation to increased temperature and Factor 2 was "Prey Treatment" with respect to non-acclimation/acclimation to increased temperature. Therefore, treatments were: "Predator Acclimated" (prey unacclimated), "Prey Acclimated" (predator unacclimated), "Both Acclimated", "Both Unacclimated". An "Ambient" temperature treatment at  $11 \pm 0.09$  °C was included to get a baseline comparison. This was the temperature of Strangford Lough during the trial period, it represents the current spring average (AFBI 2016). This was preferred to a winter average due to the very low temperatures in winter slowing feeding behaviour in preliminary trials. Lesser-spotted dogfish were selected at random, measured and placed in arenas an hour before the trials started. Recovery time was short as the specimens were captive bred and somewhat used to handling. Densities of prey were supplied at 1, 3, 5, 10, 15, 20 ( $n = 4$  per density); prey size was ~5–10 mm. Trials lasted for 2 h and the number of prey consumed were recorded at the end of this time by an observer. Individual fish were then placed in ambient temperature holding tanks for a week before reuse, no fish was reused in the same treatment or at the same prey density to avoid pseudo-replication. Amphipods were kept in experimental arenas in the absence of predators at each density treatment ( $n = 3$  per density).

## Statistical analyses

Data were non-parametrically bootstrapped ( $n = 30$ ) to calculate multiple estimates of attack rate, handling time and maximum feeding rate. Welch's t-test was used to test differences in bootstrapped parameters between "Ambient" and "Both acclimated" (i.e. "raised") groups. Generalised linear models (GLM) with quasi-

Poisson error distribution to account for over dispersion in the model were used to test for differences in functional response parameters [attack rate ( $a$ ), handling time ( $h$ ), and maximum feeding rate ( $1/hT$ )] with respect to two factors; “Predator Treatment” and “Prey Treatment”. Tukey honest significant difference post-hoc tests were used for multiple pairwise comparisons.

Functional response analyses were carried out in R v.3.3.0 (R Development Core Team 2015). The R package ‘frair’ (Pritchard 2014; Pritchard et al. 2017) was used to model the functional response Type of *Scyliorhinus canicula* preying on *Echinogammarus marinus*. Functional response Type (I, II, III) was determined using a logistic regression of the proportion of prey consumed against prey density. If the proportion of prey consumed decreases with increasing prey density, then the logistic regression will produce a significantly negative first order term and thus the functional response Type will be classed as Type II (Juliano 2001). If the logistic regression produced a non-significant first order term, a GLM was run on the proportional data, in this case a Type II response would be signified by a significantly negative first order term followed by a significantly positive second order term (Juliano 2001), while a Type III is signified by a significantly positive first order term and significantly negative second order term (Juliano 2001). A Type I functional response could be determined from a non-significant first or second order term, however a Type I is categorical only in filter feeders or systems that meet the requirements stated in Jeschke et al. (2004). Therefore, when non-significant terms were encountered, a scatterplot with locally weighted scatterplot smoothing (LOWESS) lines, with a smoothing factor of 9/10, were fitted to assess the direction of proportional consumption. This was used to assess the shape of proportional consumption of prey at different prey densities (Pritchard et al. 2017).

When a Type II response was indicated, the functional responses were modeled using Rogers’ (1972) Random Predator Equation, due to the prey not being replaced as they were consumed (Juliano 2001):

$$N_e = N_o(1 - \exp(a(N_e h - T))) \quad (1)$$

where  $N_e$  is the number of prey eaten,  $N_o$  is the initial density of prey,  $a$  is the attack rate,  $h$  is the handling time and  $T$  is the total time available. The Lambert W function was implemented to fit the model to the data (Bolker 2008).

When Type III functional responses were indicated, Hassell’s Type III response was modeled to the data wherein  $a$  is a hyperbolic function of  $N_o$  (Juliano 2001), where  $b$ ,  $c$  and  $d$  are constants and prey are depleted:

$$N_e = N_o \{1 - \exp[(d + bN_o)(hN_e - T)/(1 + cN_o)]\} \quad (2)$$

Flexible models were also modeled on the data as categorical Type II functional responses are  $q = 0$  and, as  $q > 0$ , the functional response becomes proportionally more sigmoid, with  $q = 1$  indicating a Type III response. Treatments with GLMs indicating Type III were also modeled with a flexible model with  $q$  fixed at 1 and Type II were modeled with  $q$  fixed at 0. The flexible model is an alteration of Rogers’ Random Predator Equation (Eq. 1) with the addition of a scaling exponent  $q$  (Real 1977; Vucic-Pestic et al. 2010).

$$N_e = N_o(1 - \exp(bN_o^q(hN_e - T))) \quad (3)$$

where  $N_e$  is the number of prey consumed,  $N_o$  is the initial prey density,  $b$  is the search coefficient,  $h$  is handling time,  $q$  is the scaling coefficient, and  $T$  is the total time available. The search coefficient in Eq. 3,  $b$  combines with  $N_o$  and  $q$  to produce attack rates that vary with prey density, (i.e.  $a = bN_o^q$ ) unlike in Eq. 1 where attack rate is fixed at all prey densities ( $N_o$ ).

Suitability of each model was compared with Akaike Information Criterion (AIC). Functional responses were then fitted using maximum likelihood estimation (MLE; Bolker 2010) and the Lambert W function (Bolker 2008).

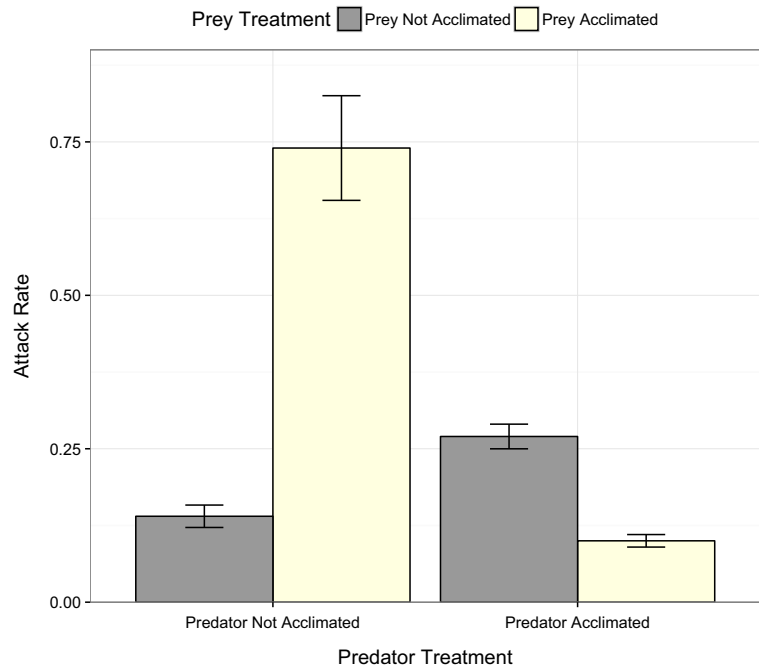
## Results

Control group prey survival was >99% and thus prey mortality was assumed to be due to predation in the experimental groups. Attack rate was significantly higher at ambient ( $0.92 \pm 0.05$  SE) than raised temperature ( $0.10 \pm 0.010$  SE) ( $t = 13.82$ ,  $df = 30.59$ ,  $p < 0.001$ ). Handling time was significantly shorter at raised temperature ( $0.06 \pm 0.008$  SE) compared to ambient ( $0.10 \pm 0.008$  SE) ( $t = 4.18$ ,  $df = 51.03$ ,  $p < 0.001$ ). This translated into a significantly higher maximum feeding estimate at raised temperature ( $29.42 \pm 4.23$  SE) compared to ambient ( $12.37 \pm 1.67$  SE) ( $t = 3.74$ ,  $df = 30.91$ ,  $p < 0.001$ ).

Attack rate had a significant “predator treatment” x “prey treatment” interaction effect ( $F_{1, 102} = 98.69$ ,  $p < 0.001$ ; Fig. 1). Wherein, predators that were non-acclimated showed higher attack rates to prey that were



**Fig. 1** Mean  $\pm$  SE of bootstrapped attack rate ( $\alpha$ ) showing the interaction of predator and prey treatments



acclimated as compared to prey that were non-acclimated, whereas predators that were acclimated showed higher attack rates toward non-acclimated as compared to acclimated prey (Fig. 1).

Handling time had a significant “predator treatment” x “prey treatment” interaction effect ( $F_{1, 102} = 32.14$ ,  $p < 0.001$ ; Fig. 2). That is, predators that were non-acclimated showed higher handling time to prey that were acclimated as opposed to non-acclimated, and predators that were acclimated showed higher handling times to prey that were non-acclimated as opposed to acclimated (Fig. 2).

Maximum feeding rate had a significant “predator treatment” x “prey treatment” interaction effect ( $F_{1, 102} = 17.75$ ,  $p < 0.00$ ; Fig. 3). When predators were non-acclimated the maximum feeding rate was similar towards prey that were non-acclimated and prey that were acclimated, whereas when predators were acclimated the maximum feeding rate was higher towards prey that were acclimated as compared to non-acclimated prey (Fig. 3).

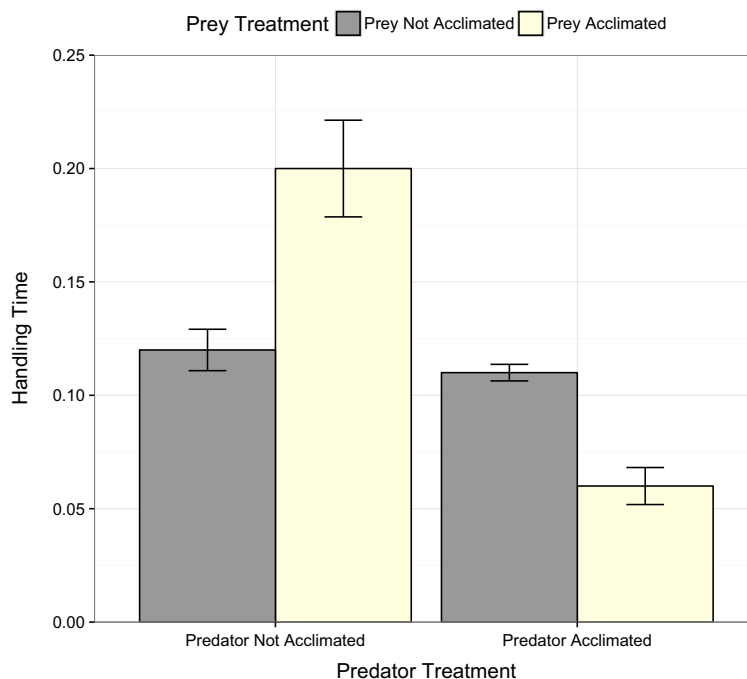
At ambient temperature (11 °C), the predator showed a Type II functional response (Table 1, Fig. 4). When just the predator was acclimated to 16.3 °C, there was no statistically significant Type III functional response, indicating that in this treatment there is a boundary Type I/Type III response. Therefore, while fitting the model of a Type I response at the densities investigated, it was

modeled as a Type III (Jeschke et al. 2004) (Table 1, Fig. 4). When both predator and prey were acclimated to 16.3 °C a Type III response was fitted to the data (Table 1, Fig. 4). This had the highest maximum feeding rate estimate. When just the prey was acclimated there was a Type II functional response (Table 1, Fig. 4). When both predator and prey were unacclimated to 16.3 °C there was a non-significant Type III functional response which, again, was instead modeled as a Type III as suggested by the shape of the GLM (Table 1, Fig. 4). This has the lowest maximum feeding rate estimate. In both instances where there was a non-significant logistic regression and GLM, locally weighted scatterplot smoothing was used on the proportion of prey consumed per density to investigate the shape of the response (Fig. 5). It should be considered here that the handling times reported for predator acclimated and both unacclimated are unreliable due to the Type I response generated at the prey densities investigated.

## Discussion

Prediction of the ecological ramifications of facets of climate change is particularly difficult due to inherent confounds of interactions between dissolved oxygen concentration and temperature (Peck et al. 2009; Grigaltchik

**Fig. 2** Mean  $\pm$  SE of bootstrapped handling time ( $h$ ) showing the interaction of predator and prey treatments

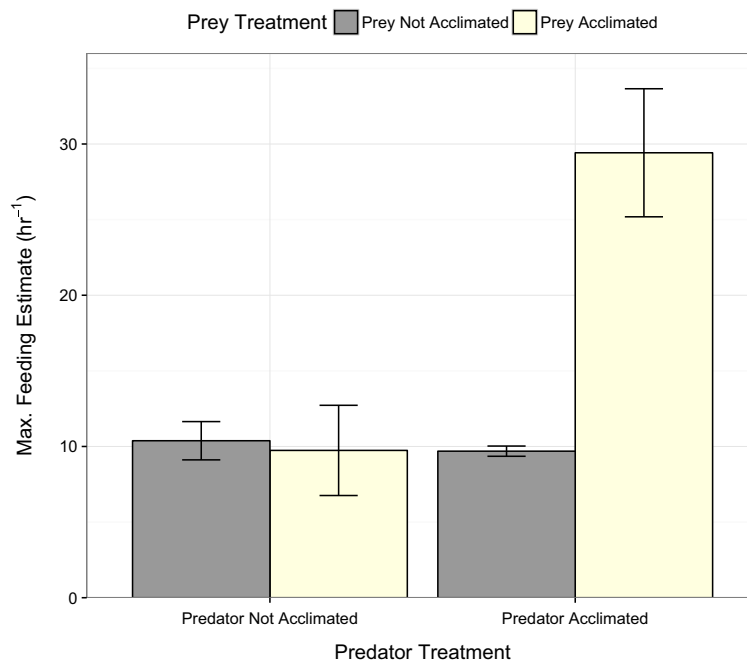


et al. 2012) and multi-species systems (Broitman et al. 2009). However, here we have shown that by using comparative functional response analyses, in a similar way to invasive species impact predictions (Dick et al. 2014; Wasserman et al. 2016b), it is possible to identify the impact of a predator under various warming scenarios

and assess whether there are interaction effects when both sides of the predator-prey dynamic are considered.

Temperature increase mediated changes to the parameters associated with the functional response. There was a significantly higher attack rate when at “Ambient” temperature compared to “Both Predator

**Fig. 3** Mean  $\pm$  SE of bootstrapped maximum feeding estimate ( $1/hT$ ) showing the interaction of predator and prey treatments



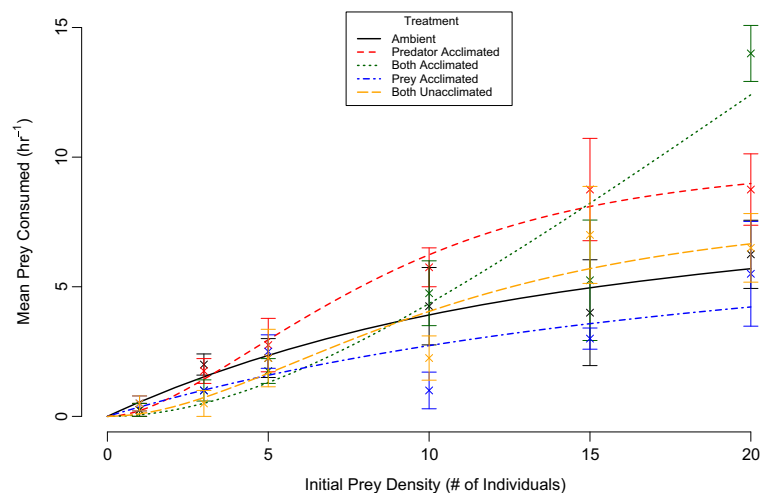
**Table 1** Logistic regression first order terms, GLM first and second order terms with associated *p* values and Functional Response Type for all experimental treatments

Treatment	(Log. Reg.) First order term, ( <i>p</i> )	(GLM) First order term, ( <i>p</i> )	(GLM) Second order term ( <i>p</i> )	Functional Response type
Ambient	−0.053, (<0.05)	−0.133, (0.25)	0.003, (0.47)	II
Predator acclimated	−0.024, (0.30)	0.164, (0.14)	−0.007, (0.08)	III
Both acclimated	0.087, (<0.001)	−0.127(<0.001)	0.008, (0.05)	III
Prey acclimated	−0.025, (0.33)	−0.365, (<0.01)	0.014, (<0.01)	II
Both unacclimated	0.008, (0.74)	0.071, (0.54)	−0.002, (0.58)	III

and Prey Acclimated” to 16.3 °C. This contrasts with the meta-analysis by Rall et al. (2012) and the assumption that attack rate is temperature invariant (Dell et al. 2014), but is similar to results in Grigaltchik et al. (2012), wherein there was a decreased attack rate but a high predation pressure at increased temperature. Further, our results are similar to Wasserman et al. (2016b), wherein attack rate was highest in the lowest temperature at zero habitat complexity. Furthermore, attack rate considers both predator and prey responses thus, at a lower temperature, swimming speed may be decreased in the smaller bodied prey species and therefore make them easier to attack as they must use more energy to escape (Dell et al. 2011; Dell et al. 2014). As a consequence, the attack rate of the predator may be higher at lower temperature due to a change in attack success as a result of prey escape energy. The handling time was significantly shorter at the higher temperature and therefore the maximum feeding rate estimates were significantly higher at the higher temperature.

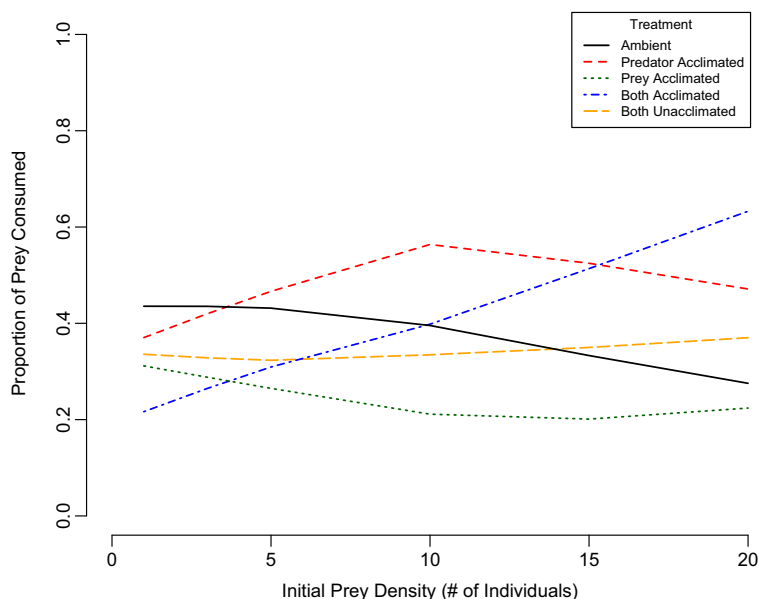
Temperature directly affects poikilotherm metabolism and has a somewhat bell shaped relationship with maximum consumption (Englund et al. 2011). At ambient temperature, dogfish take 2–3 days to fully digest a meal due to the slower rate of chemical reactions at low body temperature (Van Slyke and White 1911), thus the increase in digestion rate is comprehensible at a higher temperature (Gilbert et al. 2014). Juvenile lesser-spotted dogfish have almost double the metabolism rate of adult specimens (Sims 1996), so despite the short trial time, the predator may experience more motivational states during the raised temperature treatments due to the increased rate of digestion. This may alter searching behaviour at high prey densities (Clark and Mangel 2000; Ariyomo and Watt 2015). The behaviours that make up handling time are considered to be composite of a multitude of processes including; subduing, chewing, swallowing, digestion and gut evacuation, all of which are controlled by physiological processes that are affected in different manners by

**Fig. 4** Functional response curves of *Scyliorhinus canicula* on *Echinogammarus marinus* for all treatments with standard error bars from raw data





**Fig. 5** LOWESS lines fitted to the proportion of prey consumed at each density per treatment



increasing temperatures (Jeschke et al. 2002). This produces a complex picture, wherein, the attack rate is high at low prey densities under current temperatures which could manifest as destabilising for prey populations. Warming may mitigate the risk for low density prey populations due to the change in Functional Response Type to the stabilising Type III under raised temperatures. Nonetheless, there was a higher maximum feeding rate at the raised temperature treatment which suggests that despite a change in resource consumption with regards to resource density, the end result remains that consumption increases due to temperature increase as a result of shorter handling time.

Attack rate signifies the instantaneous rate of successful search (Holling 1959) and determines the initial increase of consumption at low resource density. When the predator was non-acclimated the attack rate was higher towards acclimated prey. This may be due to the temperature rise increasing the metabolism of the prey and causing a consequent increase in prey velocity and increased boldness (van Baalen et al. 2001; Nowicki et al. 2012). As a result, more stimuli are provided to the predator and the encounter rate is increased, leading to a higher attack rate when the prey is acclimated (Curio 1976; Taylor 1984). However, when the predator was acclimated the attack rate was higher toward non-acclimated prey and lower toward prey that were acclimated. This suggests that acclimation to temperature may dampen attack rate in juvenile lesser-spotted dogfish and thus negate the effects of increased prey stimuli.

Handling time relates to the time taken to eat and digest prey items (Jeschke et al. 2002). Unfortunately, we cannot discuss the handling times of the treatments “Predator Acclimated”, and “Both Unacclimated” with conviction, due to the non-significant logistic regression and GLM first and second order terms. Nonetheless, we can speculate from the results accumulated that at higher prey densities there may be further differences in resource consumption and, thus, handling times, as the motivational state changes (Englund et al. 2011). When the predator was non-acclimated the handling time was longer towards prey that were acclimated. This demonstrates that prey acclimation to temperature has a large effect on the handling of the prey possibly due to a mismatch in swimming speed of predator and the prey, and digestion speed in the predator (Dell et al. 2011). When the predator was acclimated the handling time was longer towards prey that were non-acclimated. This may again be related back to encounter rate (Curio 1976; Taylor 1984) and an increased metabolism and digestion speed due to acclimation of the predator (Dell et al. 2011). Therefore, when predator and prey are both acclimated to the same temperature there is a decrease in handling time which implies that consumption rates will also increase with temperature warming. While the effects of oxygen concentration were not tested in the present study, air was bubbled through all arenas during the treatments. Oxygen is a respiratory limiting factor at high temperatures and this effect is increased during acute warming scenarios, wherein there is no

acclimation as organisms consume more oxygen when exposed to an abrupt temperature change than after an acclimation period (Bullock 1955; Kitching 1977).

Maximum feeding rate estimations are derived from functional response handling times (Juliano 2001). When the predator was non-acclimated the maximum feeding rate did not change with prey treatment but when the predator was acclimated the maximum feeding rate was far higher when prey were also acclimated. This indicates that when predator and prey populations undergo simultaneous acclimation to higher temperatures there will be a large increase in resource consumption, presumably due to the temperature mediation of the processes associated with handling time. When considered with the functional response types, it seems that simultaneous heating and acclimation produce an additive affect, which increases the maximum feeding rate and predator impact when prey is at 'high' densities. Although this must be caveated with the unreliable handling estimates of "Predator Acclimated", and "Both Unacclimated" which could be revealed under further investigation to have higher maximum feeding estimates.

There was a high variability in the shape of functional response in the different treatments with, justifiably, all three categorical functional responses observed. However, Type I functional responses require simultaneous handling and digestion and thus are only reported definitively amongst filter feeders (Lehman 1976; Jeschke et al. 2002; Jeschke et al. 2004). The intermediate response that was modeled is likely due to the low densities of prey supplied during the treatments (Jeschke et al. 2004), or the lack of habitat complexity in the experimental arena reducing handling limitations. As such, the results indicating Type I boundary responses may not directly reflect feeding ecology in the wild, but in this case serve as a comparison between the relative differences in parameter values. The results obtained were modeled as Type III responses due to the difference in proportional consumption. A Type III functional response was prevalent in trials at temperature apart from "Prey Acclimated". The change from Type II to Type III is concordant with other studies using fish predators which report shifts with environmental changes (Koski and Johnson 2002; Alexander et al. 2012), but contrasts with Wasserman et al. (2016b), which observed no functional response change in an invertebrate predator – prey system over various temperatures and habitat complexity. The Type III response when both were acclimated may be accounted for by the increased

movement at high temperature (Biro et al. 2010) causing low consumption at low resource densities, but as the prey density increases there is a higher encounter rate and thus higher consumption. Furthermore, it is considered that when mortality is independent of foraging, foraging activity should increase concomitantly with resource availability (McNamara and Houston 1994). This would be the case when related to *S. canicula* in the Irish Sea where they are under low predation pressure. "Both Acclimated" had the highest magnitude functional response, while "Prey Acclimated" had the lowest. As "Prey Acclimated" functional response intensity was less than "Ambient" it suggests that the effect of temperature acclimation on *E. marinus* is more profound than on the predator, potentially because of the small body size of *E. marinus* giving the prey a metabolic advantage as the increased metabolic demands are mitigated by the small body size compared to the larger predator (Lincoln 1979; Whiteley et al. 2011; Reuman et al. 2014; Di Santo and Lobel 2016).

The combined additive effects of predator and prey acclimation appear to increase predator impact, while temperature increase drives a shorter handling time despite reducing the attack rate. While the present study deals specifically with juvenile lesser-spotted dogfish the results remain informative of the changes that temperature exerts on predation. This could have implications on the impacts of wild populations of *S. canicula* as they show spatial segregation by sex, temporal segregation and varying thermal strategies (Sims et al. 2001) as this may mediate the predation impacts when moving between thermally varied micro-climates. It is considered that when functional response is delivered as a phenomenological tool, the focus shifts from the shape of the response to the relative differences in parameter estimates (Lehman 1976; Jeschke et al. 2002; Alexander et al. 2012). This is due, in part, to the laboratory set up being inherently dissimilar to the conditions in situ with regards to variation in temperature, prey availability, arena size, lack of refugia and habitat complexity. Nonetheless, we have provided a useful parameterization of the differences in resource consumption when predator and prey are factorially exposed to temperature increase.

This study demonstrates the importance of temperature as a regulating driver of predation. Temperature has the ability to increase predation pressure. Due to the high densities of *Scylliorhinus canicula* (Richardson 2016), increasing population size (Richardson 2016), and high discard survival rate (Revill et al. 2005), there is

likelihood that predation on important fisheries species will increase. However, due to the generalist feeding strategy (Murdoch 1969) and patchy spatial distribution of prey, the Type III response is likely to prevail in the field (Trenkel et al. 2004).

Furthermore, *Echinogammarus marinus* may have an evolutionary advantage, due to body size, at higher temperatures (Genner et al. 2010; Di Santo and Lobel 2016), yet this may be mitigated when prey is found in high densities. The rate of warming, and subsequently environmental variation, affects maximum feeding estimates; consequently variation from the mean temperature is likely to have an impact upon predation pressure (Dell et al. 2011; Paaijmans et al. 2013; Vasseur et al. 2014). Furthermore, patchiness of prey populations could confer a benefit against predation, despite increased predation pressured conferred through decreased handling time. Though, it is necessary to caveat with the emphasis that thermal responses are species specific (Helmuth et al. 2002; Broitman et al. 2009; Englund et al. 2011), as is the “environment” that each species experiences (Kearney 2006; Broitman et al. 2009), but also highly variable between individuals (Spicer and Gaston 1999; Calosi et al. 2013).

It is important to understand and quantify interaction strength in food webs and the how changes in abiotic context will affect predator-prey dynamics with regards to species and individual levels within a system (Calosi et al. 2013). The functional responses and associated parameters shown here demonstrate the differential responses within two organisms to a range of climatic scenarios. This will have implications for the sustainability of fisheries in the Irish and Celtic Seas, wherein there is high abundance of *Scyllorhinus canicula*, as acclimation to increased temperature increases maximum feeding rate but changes the proportional consumption of sparse prey populations.

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#### Compliance with ethical standards

**Ethics statement** All applicable international, national and/or institutional guidelines for the care and use of animals were followed.

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